

# An A<sub>2</sub>O-MBR system for simultaneous nitrogen and phosphorus removal from brewery wastewater

Van Nu Thai Thien<sup>1</sup>, Dang Viet Hung<sup>2,\*</sup>, Nguyen Thi Thanh Hoa<sup>3</sup>

## ABSTRACT

Anaerobic/Anoxic/Oxic – Membrane BioReactor (A<sub>2</sub>O-MBR) system was used to enhance simultaneous removal of nitrogen and phosphorus from brewery wastewater. The A<sub>2</sub>O unit containing microorganisms with short solids retention time (SRT) was employed mainly for removal of organic matter and phosphorus together with denitrification. The MBR containing microorganisms with long SRT was employed mainly for nitrification of NH<sub>4</sub><sup>+</sup>-N and recirculation of NO<sub>3</sub><sup>-</sup>-N. The model of A<sub>2</sub>O-MBR system made from polyacrylic with the capacity of 49.5 liters was operated with hydraulic retention times decreased from 24, 18 to 12 hours corresponding to organic loading rates increased from 0.50, 0.75 to 1.00 kg COD/m<sup>3</sup>.day. The results showed that the model not only treated organic matter well but also nearly completely removed both nitrogen and phosphorus. For all three loading rates, chemical oxygen demand (COD) concentration decreased significantly in the anaerobic and anoxic compartments of the A<sub>2</sub>O unit, indicating that most of organic matter was utilized in the anaerobic and anoxic compartments for phosphorus release and denitrification, respectively. Nitrification in the MBR was almost perfectly completed, with average NH<sub>4</sub><sup>+</sup>-N removal efficiencies of over 98%. Denitrification in the anoxic compartment happened as much as possible. Demands for the development of PAOs, which were responsible for enhanced biological phosphorus removal (EBPR) processes, could be provided. For loading rate of 0.75 kg COD/m<sup>3</sup>.day, treatment efficiencies of COD, NH<sub>4</sub><sup>+</sup>-N, total nitrogen (TN) and total phosphorus (TP) of the model were the highest as 95.4, 99.2, 86.7 and 84.6%, respectively. Output values of these parameters were within the limits of Vietnam National Technical Regulation on Industrial Wastewater (QCVN 40:2011/BTNMT), column A. The model of A<sub>2</sub>O-MBR system was capable of achieving effluents with very low nitrogen and phosphorus concentrations from brewery wastewater.

**Key words:** A<sub>2</sub>O-MBR system, brewery wastewater, nitrogen removal, phosphorus removal

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## History

- Received: 18-01-2019
- Accepted: 14-5-2019
- Published: 20-6-2019

## DOI :

<https://doi.org/10.32508/stdjsee.v3i1.507>



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## INTRODUCTION

By Vietnam Beer Alcohol Beverage Association, Vietnamese people consumed nearly 4.1 billion liters of beer in 2017. Currently, there are approximately 129 brewery production facilities across the country with the installed capacity of 4.8 billion litres of beer. Along with this consumption, serious problems with environmental pollution may be caused by a huge amount of brewery wastewater. This amount of wastewater must be treated before discharge into environment. To brewery wastewater, a combination anaerobic-aerobic treatment system has been used and traditional aerobic biological treatment processes such as activated sludge (suspended growth) or biological filter (attached growth) are often implemented<sup>1-4</sup>. However, these processes have not yet treated thoroughly nitrogen and phosphorus from brewery wastewater to meet QCVN 40:2011/BTNMT, column A.

Anaerobic/Anoxic/Oxic (A<sub>2</sub>O) process commonly used in wastewater treatment is able to remove or-

ganic matter together with nitrogen and phosphorus with its own inherent advantages such as short hydraulic retention time (HRT), high pollutant removal efficiency and good shock loading capacity<sup>5,6</sup>. The process consists of three anaerobic, anoxic, oxic compartments and one settling tank which are arranged in sequence with nitrate circulating flow from the oxic compartment to the anoxic compartment and sludge circulating flow from the settling tank to the anaerobic compartment. In this process, nitrification by nitrifiers occurs in the oxic compartment; denitrification by denitrifiers in the anoxic compartment; absorption of  $\beta$ -polyhydroxybutyrate (PHB) for phosphate release by Phosphorus Accumulating Organisms (PAOs) in the anaerobic compartment and then oxidation of PHB for phosphorus accumulation in the oxic compartment. Excess sludge discharge occurs in the settling tank<sup>7</sup>.

However, A<sub>2</sub>O process is a single sludge process with the only line for excess sludge discharge at the settling tank so there has been limitation to satisfy a proper

**Cite this article :** Thai Thien V N, Viet Hung D, Thanh Hoa N T. **An A<sub>2</sub>O-MBR system for simultaneous nitrogen and phosphorus removal from brewery wastewater.** *Sci. Tech. Dev. J. - Sci. Earth Environ.*; 3(1):12-22.

SRT for both nitrifiers and PAOs in the oxic compartment of A<sub>2</sub>O process<sup>8,9</sup>. On the other hand, nitrifiers need long SRT and PAOs need short SRT. To solve this problem, incorporation of a biological reactor into A<sub>2</sub>O unit, so-called A<sub>2</sub>O – Biological Reactor system, for simultaneous nitrogen and phosphorus removal has been attempted in the past decade<sup>8-10</sup>. The A<sub>2</sub>O unit containing microorganisms with short SRT is employed mainly for removal of organic matter and phosphorus together with denitrification. The biological reactor containing microorganisms with long SRT is employed mainly for nitrification of NH<sub>4</sub><sup>+</sup>-N and recirculation of NO<sub>3</sub><sup>-</sup>-N.

Weitang Zhang *et al.* (2013) studied removal of nutrient from domestic wastewater with low COD/N ratio by an A<sub>2</sub>O–Biological Aerated Filter (A<sub>2</sub>O-BAF) system. The favorable  $V_{anoxic} / V_{oxic}$  was from 2.5:1 to 6:1. When  $V_{anoxic} / V_{oxic}$  was 6:1, treatment efficiencies of COD, TN and PO<sub>4</sub><sup>3-</sup>-P achieved very high values as 89 ± 4, 83 ± 3 and 99 ± 1%, respectively<sup>11</sup>. Recently, membrane bioreactor (MBR) is an attractive process that has been increasingly used for advanced biological wastewater treatment. With membrane filtration replacing secondary clarification, MBR possesses a number of merits such as biomass enrichment, perfect nitrification, small footprint, ensured sludge-effluent separation, easy manipulation of HRT and SRT, and excellent effluent quality with little organic and solid contents<sup>12-15</sup>. Thus, MBR was selected as Biological Reactor in the combined system because of the capacity to achieve enhanced nitrification rate and produce high quality effluent<sup>16,17</sup>.

In this study, an A<sub>2</sub>O-MBR system was used to evaluate the effects of loading rate on the combined system's simultaneous nitrogen and phosphorus removal performance via continuous flow by treating real brewery wastewater. The role of MBR in the combined system and its contribution to organic matter, nitrogen and phosphorus removal were also investigated.

## MATERIALS AND METHODS

### Experimental

The polyacrylic model of A<sub>2</sub>O-MBR system included Anaerobic/Anoxic/Oxic unit having an approximate dimension of 480 mm L x 150 mm W x 600 mm H with the corresponding working volume of 36.0 liters which was divided by baffles to create three compartments (anaerobic, anoxic, oxic) in ratio of 2:4:2<sup>11</sup> and MBR having an approximate dimension of 180 mm L x 150 mm W x 600 mm H with the corresponding working volume of 13.5 liters. Total working volume of the model was 49.5 liters. Settling tank had

an approximate dimension of 150 mm D x 300 mm H with the working volume of 7.2 liters. In the MBR, a polyethylene hollow-fiber membrane module (0.4 μm pore size, 0.32 m<sup>2</sup> effective area, Mitsubishi Rayon Co., Ltd, Japan) was immersed. Aeration was provided through fine air diffusers from the bottoms in the oxic compartment and MBR while sludge in the anaerobic and anoxic compartments was suspended by paddle mixers at 50 rpm. Effluent was withdrawn through the membrane module by a suction pump that was designed for intermittent operation with a duty cycle of 8 minutes ON / 2 minutes OFF. To mitigate membrane fouling, backflushing was carried out every 24 hours for 15 min. Dissolved oxygen (DO) concentrations of the oxic compartment and MBR were determined by DO meter and controlled from 2 to 4 mg/L<sup>18</sup>. Return effluent ratio of 200% and return sludge ratio of 100% were fixed<sup>9</sup>. Schematic representation of the experimental system was represented in Figure 1.

- 1/Wastewater tank: 200 liters (PE, Vietnam);
- 2/Anaerobic/Anoxic/Oxic unit with three compartments: 36.0 liters (Polyacrylic, Vietnam);
- 3/Settling tank: 7.2 liters (Polyacrylic, Vietnam);
- 4/MBR with a polyethylene hollow-fiber membrane module: 13.5 liters (Polyacrylic, Vietnam);
- 5/Middle tank: 50 liters (PE, Vietnam);
- 6/Feed pump: 11 liters/hour (Sandur, India);
- 7/Effluent pump: 16 liters/hour (Sandur, India);
- 8/Suction pump: 11 liters/hour (Blue & White, United State);
- 9/Sludge pump: 11 liters/hour (Sandur, India);
- 10/Paddle mixer 1: (IWAKI, Japan);
- 11/Paddle mixer 2: (IWAKI, Japan);
- 12/Blower 1: 38 liters/min (RESUN, Ap 001, China);
- 13/Blower 2: 38 liters/min (RESUN, Ap 001, China);
- 14/Sludge valve 1: Ø13 (Copper, Vietnam);
- 15/Sludge valve 2: Ø13 (Copper, Vietnam).

### System operating conditions

The wastewater treatment experiment was conducted in four phases in the laboratory at room temperature (~ 25°C). In the short initial phase, so-called phase 0, seed sludge was given to 50% volume of the model with MLSS concentration about 5000 mg/L. Influent wastewater with average COD concentration of 500 mg/L diluted with tap water was pumped into the model. Organic loading rate was increased little by little from 0.1 to 0.3 kgCOD/m<sup>3</sup>.day. The phase 0 ended when COD removal efficiency remained stable at above 80%. There was no sludge discharged except sampling to keep large amounts of biomass.

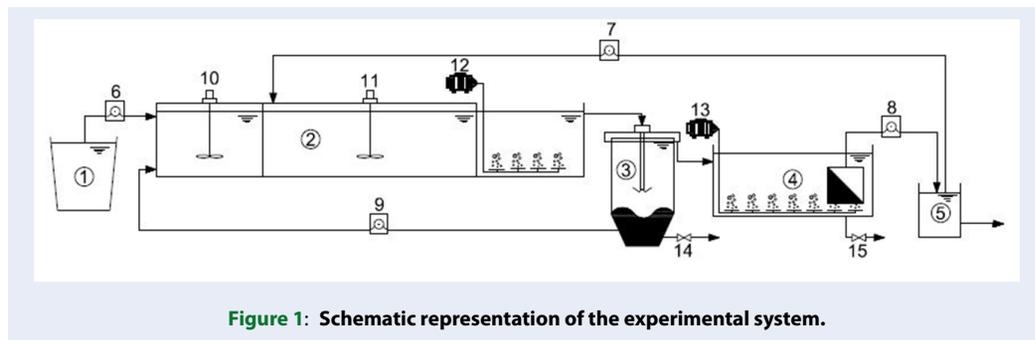


Figure 1: Schematic representation of the experimental system.

In the next three phases according to overall treatment performance in relation to the different loading rates, denoted as 1, 2 and 3, respectively, raw wastewater was pumped continuously with wastewater flow rates increased from 49.5 to 99.0 liters/day corresponding to HRTs decreased from 24 to 12 hours and organic, nitrogen, phosphorus loading rates increased from 0.5 to 1.0 kgCOD/m<sup>3</sup>.day, 0.08 to 0.16 kgTN/m<sup>3</sup>.day, 0.014 to 0.028 kgTP/m<sup>3</sup>.day, respectively as in Table 1. Excess sludge was discharged from the A<sub>2</sub>O unit and MBR to maintain SRTs from 5 to 7 days and from 45 to 60 days, respectively.

Trans-membrane pressure (TMP) was used as an indicator of membrane fouling and monitored continuously by a data logging manometer. When TMP reached 40 kPa, membrane washing was performed physically and chemically following the guidelines of the manufacturer. In the phases 0, 1, 2 and 3, the membrane module was physically washed on a daily basis for 15 min. During the entire period of experiment, the TMP was maintained below 40 kPa. Therefore, the membrane module was not cleaned chemically.

### Wastewater source

Brewery wastewater came from the outlet of the UASB reactor of Wastewater Treatment Plant at Nguyen Chi Thanh – Saigon Beer Manufacturing Factory, Ho Chi Minh City, Vietnam. The main characteristics of influent wastewater were presented in Table 1. Seed sludge for the model of A<sub>2</sub>O-MBR system was taken from one of the two SBRs of this wastewater treatment plant. Seed sludge was light brown, well-settled with sludge volume index of 98 and MLVSS/MLSS ratio of 0.74.

### Analytical methods

The samples were collected at the input and output positions of the experimental system. They were also collected in three compartments of the A<sub>2</sub>O unit.

For each loading rate, the model was operated for 45 days to achieve a steady-state condition and the samples were collected over a 3-day period during these days. For determination of the overall treatment performance in terms of organic and nutrient removals, the parameters of wastewater such as COD, suspended solid (SS), Total Kjeldahl Nitrogen (TKN), NH<sub>4</sub><sup>+</sup>-N, NO<sub>2</sub><sup>-</sup>-N, NO<sub>3</sub><sup>-</sup>-N and TP were analyzed according to Vietnam National Standards together with Standard Methods for the Examination of Water and Wastewater (APHA, AWWA, and WEF)<sup>19</sup> at Research Institute for Aquaculture No.2 in Ho Chi Minh City. The value for TN was based on the sum of TKN, NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N. pH and DO were measured by pH (Mettler Toledo MP220, Switzerland) and DO (YSI 5000, United States) meters, respectively. The results below were based on average value and standard deviation by using Microsoft Office Excel software.

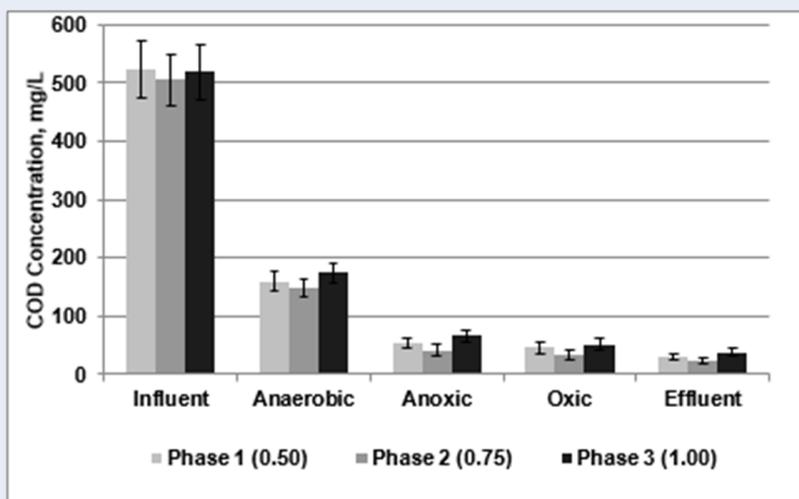
## RESULTS -DISCUSSION

### Organic removal efficiency

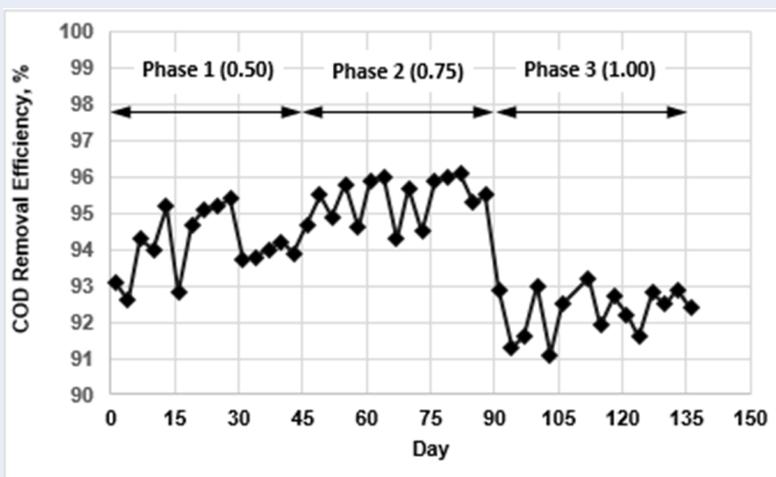
COD concentrations at different positions in the model were revealed in Figure 2 for loading rates of 0.50, 0.75 and 1.00 kgCOD/m<sup>3</sup>.day. The results showed that COD concentration decreased significantly in the anaerobic and anoxic compartments. The decline could be attributed mainly by the dilution and uptake. About 40% of COD was utilized in the anaerobic compartment by PAOs and 40% of COD was consumed in the anoxic compartment by denitrifiers<sup>10,20</sup>. It changed slightly in the oxic compartment and the MBR. The additional organic removal was attributable to the step of membrane filtration which is beneficial to keep a higher COD removal efficiency<sup>21,22</sup>. Accumulation of PO<sub>4</sub><sup>3-</sup>-P by PAOs happened mostly in the oxic compartment. Nitrification of NH<sub>4</sub><sup>+</sup>-N by nitrifiers happened mostly in the MBR. Before wastewater flowed into the MBR, large amount of COD in wastewater was removed. It

**Table 1: THE EXPERIMENTAL CONDITION IN DIFFERENT PHASES FOR THE MODEL OFA<sub>2</sub>O-MBR SYSTEM**

Phase	Duration (day)	COD (mg/L)	NH <sub>4</sub> <sup>+</sup> -N (mg/L)	TP (mg/L)	Organic loading (kgCOD/m <sup>3</sup> .day)	HRT (h)
1	1 - 45	523 ± 48	67 ± 9	13 ± 3	0.50	24
2	46 - 90	505 ± 43	70 ± 10	15 ± 3	0.75	18
3	91 - 135	518 ± 47	69 ± 10	14 ± 3	1.00	12



**Figure 2: Change of COD concentration at various loading rates.**



**Figure 3: COD removal efficiencies at various loading rates.**

was considered to be advantageous for the nitrification because of non-inhibitory effects. Therefore, the growth of nitrifiers was favourable and the nitrification was enhanced as well. COD removal efficiencies at various loading rates of the model were represented in Figure 3. For loading rates of 0.50, 0.75 and 1.00 kg COD/m<sup>3</sup>.day, average COD removal efficiencies of the model were 94.1, 95.4 and 92.3%, respectively. It could be seen that COD removal efficiency reached the highest value at the proper loading rate of 0.75 kgCOD/m<sup>3</sup>.day. For these three loading rates, output values of COD were within the limits of QCVN 40:2011/BTNMT, column A. COD removal at different loading rates depended on nitrogen and phosphorus removal mutually through treatment performance.

### Nitrogen removal efficiency

Nitrogen concentrations at different positions in the model were revealed in Figures 4 and 5 and Figure 6 for loading rates of 0.50, 0.75 and 1.00 kgCOD/m<sup>3</sup>.day, respectively. The results showed that NH<sub>4</sub><sup>+</sup>-N and TN concentrations decreased significantly in the anaerobic and anoxic compartments. The decline could be attributed mainly by the dilution of the return sludge flow in the anaerobic compartment and denitrification by denitrifiers in the anoxic compartment. It also showed that TN at the oxic compartment and MBR was mostly NH<sub>4</sub><sup>+</sup>-N and NO<sub>3</sub><sup>-</sup>-N, respectively. Due to membrane separation, a sufficiently long SRT necessary to prevent the washout of nitrifiers was applied in the MBR to improve the nitrification capability of activated sludge<sup>12</sup>. Small amount of NH<sub>4</sub><sup>+</sup>-N was metabolized for the growth of microorganisms in the system and the remaining was almost completely transformed by the nitrification in the MBR. Very low NO<sub>3</sub><sup>-</sup>-N concentration in the anoxic compartment indicated that the denitrification happened as much as possible in this compartment<sup>9</sup>. It was fully reasonable with the change of COD stated above. Removal efficiencies of nitrogen at various loading rates of the model were represented in Figure 7. For loading rates of 0.50, 0.75, 1.00 kgCOD/m<sup>3</sup>.day, average NH<sub>4</sub><sup>+</sup>-N and TN removal efficiencies of the model were 99.1 and 83.7, 99.2 and 86.7, 98.7 and 82.5%, respectively. Nitrogen removal efficiency also reached the highest values at the proper loading rate of 0.75 kg COD/m<sup>3</sup>.day. For all three loading rates, output values of NH<sub>4</sub><sup>+</sup>-N and TN were within the limits of QCVN 40:2011/BTNMT, column A. COD and nitrogen removal decreased when loading rate increased.

### Phosphorus removal efficiency

Phosphorus concentrations at different positions in the model were revealed in Figure 8 for loading rates of 0.50, 0.75 and 1.00 kgCOD/m<sup>3</sup>.day. The results showed that TP concentration increased to the maximum level in the anaerobic compartment when PAOs released phosphate by utilizing 40% of COD in wastewater as mentioned above. Conditions that favor the growth of PAOs and anaerobic phosphorus release could be provided. TP concentration decreased in the anoxic compartment by the dilution of the return effluent flow from the MBR. In addition, TP concentration also decreased significantly in the anoxic compartment due to its uptake by Denitrifying Phosphorus Accumulating Organisms (DPAOs), which could use nitrate and/or nitrite rather than oxygen as an electron acceptor when exposed to an anoxic environment. In the oxic compartment, TP was further accumulated by PAOs to reach complete biological phosphorus removal. Yongzhi Chen et al., 2011 also showed that DPAOs played an important role in removing almost entirely phosphorus from wastewater when treating domestic wastewater by an A<sub>2</sub>O-BAF system<sup>9</sup>. Phosphorus removal efficiencies at various loading rates of the model were represented in Figure 9. For loading rates of 0.50, 0.75 and 1.00 kgCOD/m<sup>3</sup>.day, average TP removal efficiencies of the model were 74.6, 84.6 and 73.5%, respectively. Phosphorus removal efficiency also reached the highest values at the proper loading rate of 0.75 kgCOD/m<sup>3</sup>.day. For all three loading rates, output values of TP were within the limits of QCVN 40:2011/BTNMT, column A. In relation to the results obtained above, the more COD removal or cell growth is, the more phosphorus removal is.

### Membrane fouling

Membrane fouling in MBR was inevitable. The TMP in the MBR of the model was monitored continuously to evaluate the membrane fouling during the entire running period. The TMP was in the range of 10 – 33 kPa and the flux was from 6.4 to 12.8 L/m<sup>2</sup>.h (LMH). The membrane fouling rate in the MBR correlates well with the MLSS concentration<sup>23</sup>. Figure 10 and Figure 11 show the variations of TMP and MLSS concentration during 135 days of operation. The MLSS concentration initially increased from around 5600 mg/L to nearly 6000 mg/L on day 60 and was maintained for the remaining days of running. When the flux was 6.4 LMH in the phase 1, the TMP was in the range of 10 – 16 kPa for 45 days. During the phase 2, the flux was kept at 9.6 LMH. The TMP increased gradually with

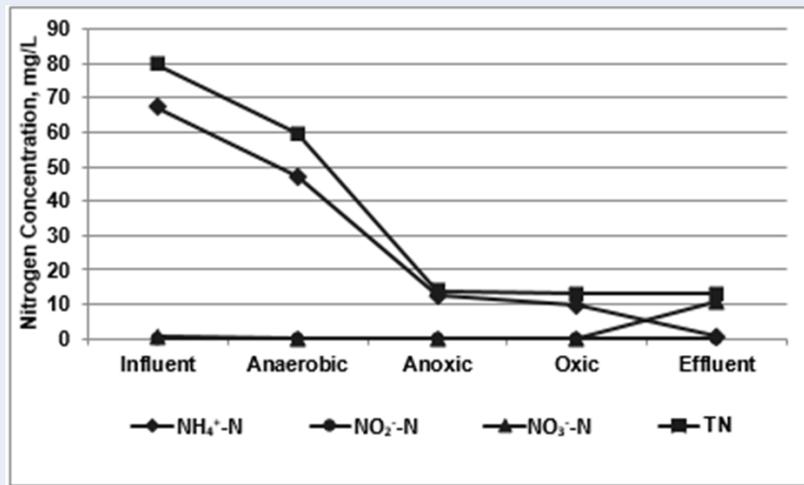


Figure 4: Conversion of nitrogen concentration for a loading rate of 0.50 kgCOD/m<sup>3</sup>.day.

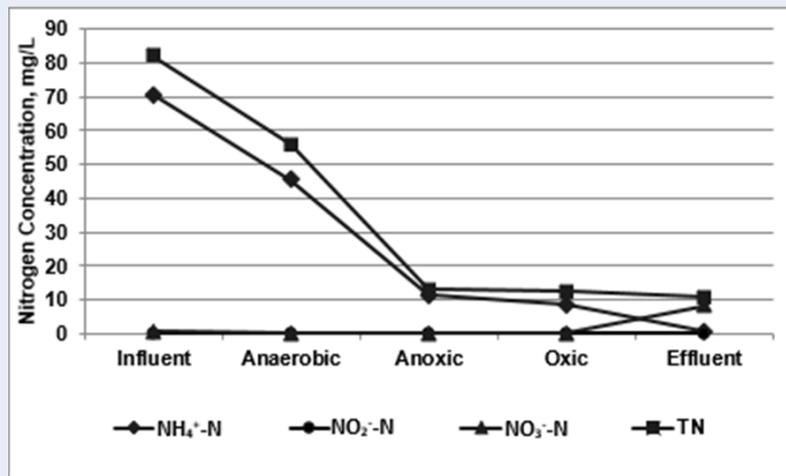


Figure 5: Conversion of nitrogen concentration for a loading rate of 0.75 kgCOD/m<sup>3</sup>.day.

time to 26 kPa on day 90. After the phase 2, the flux increased again to 12.8 LMH in the phase 3. The TMP increased almost linearly and reached about 33 kPa on day 135. As mentioned above, the membrane fouling could be alleviated to a certain degree by the intermittent operation of the membrane (2 min rest in every 10 min operation), air bubbling and backflushing.

### CONCLUSIONS

In this study, the model of A<sub>2</sub>O-MBR system was operated well and treatment efficiencies of nitrogen and phosphorus at three loading rates were high. It was capable of achieving effluents with low nitrogen and phosphorus concentrations from brewery wastewater. For a loading rate of 0.75 kg COD/m<sup>3</sup>.day, treat-

ment efficiencies of COD, NH<sub>4</sub><sup>+</sup>-N, TN, TP of the model were the highest as 95.4, 99.2, 86.7, 84.6%, respectively. Output values of these parameters were within the limits of QCVN 40:2011/BTNMT, column A. Making a short SRT for A<sub>2</sub>O unit and a long SRT for MBR helps A<sub>2</sub>O-MBR system remove simultaneously nitrogen and phosphorus from wastewater.

### LIST OF ABBREVIATIONS

- A<sub>2</sub>O-MBR: Anaerobic/Anoxic/Oxic – Membrane BioReactor
- SRT: Solids Retention Time
- COD: Chemical Oxygen Demand
- EBPR: Enhanced Biological Phosphorus Removal
- TN: Total Nitrogen

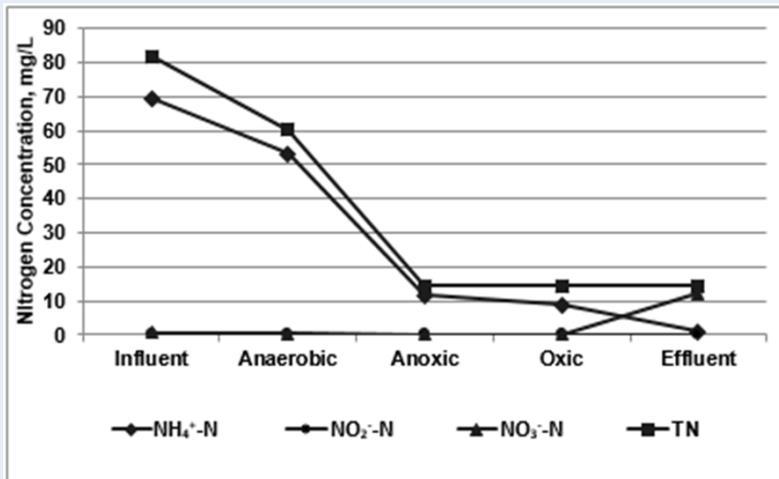


Figure 6: Conversion of nitrogen concentration for a loading rate of 1.00 kgCOD/m<sup>3</sup>.day.

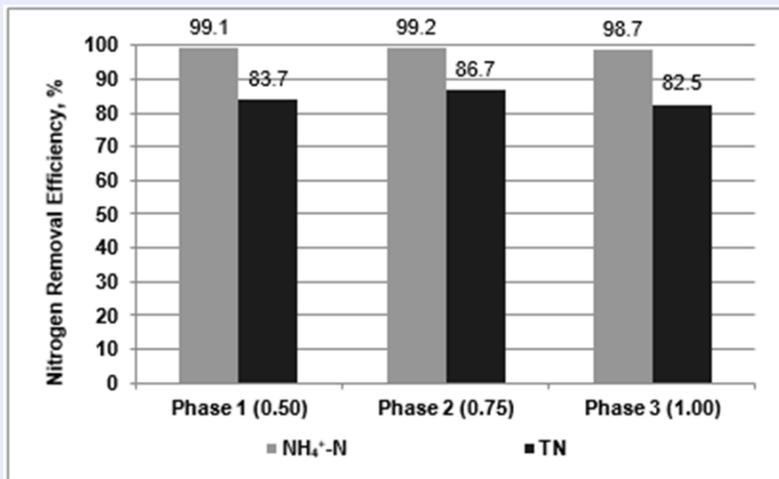


Figure 7: Nitrogen removal efficiencies at various loading rates.

TP: Total Phosphorus

QCVN 40:2011/BTNMT: Vietnam National Technical Regulation on Industrial Wastewater

HRT: Hydraulic Retention Time

PHB:  $\beta$ -polyhydroxybutyrate

PAOs: Phosphorus Accumulating Organisms

A<sub>2</sub>O-BAF: A<sub>2</sub>O-Biological Aerated Filter

DO: Dissolved Oxygen

TMP: Trans-Membrane Pressure

SS: Suspended Solid

TKN: Total Kjeldahl Nitrogen

DPAOs: Denitrifying Phosphorus Accumulating Organisms

LMH: L/m<sup>2</sup>.h

## COMPETING INTERESTS

None of the authors reported any conflict interest related to this study.

## AUTHORS' CONTRIBUTIONS

Van Nu Thai Thien: writing the draft of the research paper.

Dang Viet Hung: designing and conducting the experiments.

Nguyen Thi Thanh Hoa: sampling and analysis of wastewater.

## REFERENCES

1. Simate GS, Cluett J, Iyuke SE, Musapatika ET, Ndlovu S, Walubita LF, et al. The treatment of brewery wastewater for reuse:

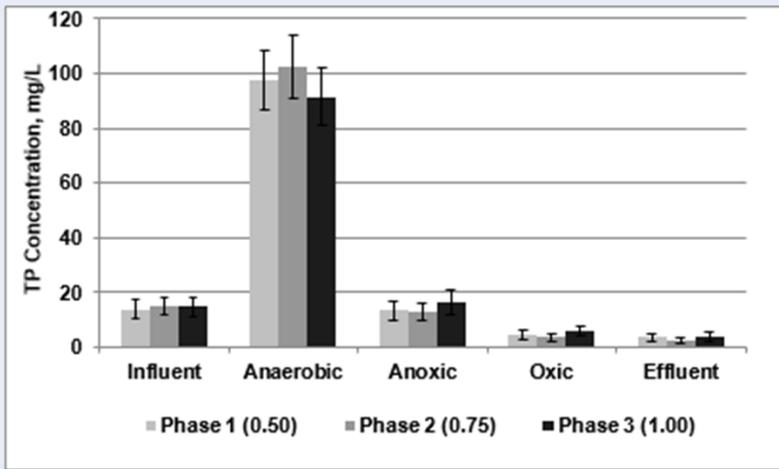


Figure 8: Conversion of TP concentration at various loading rates.

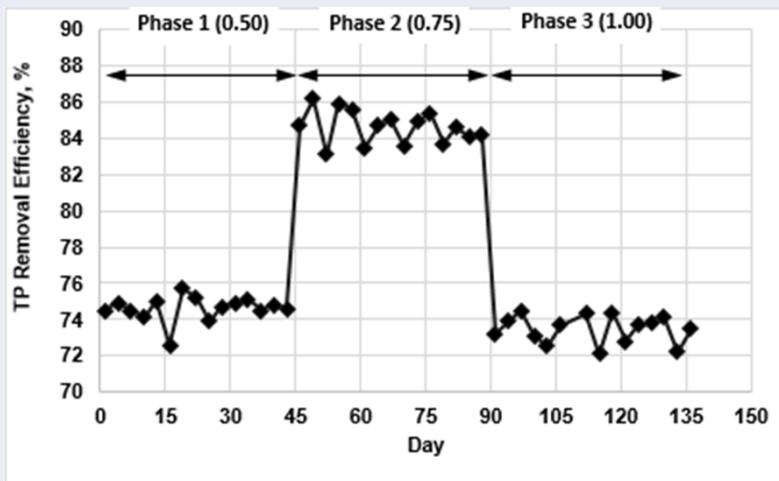


Figure 9: TP removal efficiencies at various loading rates.

State of the art. *Desalination*. 2011;273(2-3):235–247. Available from: [10.1016/j.desal.2011.02.035](https://doi.org/10.1016/j.desal.2011.02.035).

- Annie Hill. *Brewing Microbiology: Managing Microbes, Ensuring Quality and Valorising Waste*. Woodhead Publishing; 2015, p. 425–456. 1st edition. Available from: [10.1016/C2014-0-03102-4](https://doi.org/10.1016/C2014-0-03102-4).
- Bakare BF, Shabangu K, Chetty M. Brewery wastewater treatment using laboratory scale aerobic sequencing batch reactor. *South African Journal of Chemical Engineering*. 2017;24:128–134. Available from: [10.1016/j.sajce.2017.08.001](https://doi.org/10.1016/j.sajce.2017.08.001).
- Wang SG, Liu XW, Gong WX, Gao BY, Zhang DH, Yu HQ. Aerobic granulation with brewery wastewater in a sequencing batch reactor. *Bioresource Technology*. 2007;98(11):2142–2147. Available from: [10.1016/j.biortech.2006.08.018](https://doi.org/10.1016/j.biortech.2006.08.018).
- Ma Y, Peng Y, Wang X. Improving nutrient removal of the AAO process by an influent bypass flow by denitrifying phosphorus removal. *Desalination*. 2009;246(1-3):534–544. Available from: [10.1016/j.desal.2008.04.061](https://doi.org/10.1016/j.desal.2008.04.061).
- Ge S, Zhu Y, Lu C, Wang S, Peng Y. Full-scale demonstration of step feed concept for improving an anaero-

bic/anoxic/aerobic nutrient removal process. *Bioresource Technology*. 2012;120:305–313. Available from: [10.1016/j.biortech.2012.06.030](https://doi.org/10.1016/j.biortech.2012.06.030).

- Peng YZ, Wang XL, Li BK. Anoxic biological phosphorus uptake and the effect of excessive aeration on biological phosphorus removal in the A2O process. *Desalination*. 2006;189:155–164. Available from: [10.1016/j.desal.2005.06.023](https://doi.org/10.1016/j.desal.2005.06.023).
- Wang J, Peng Y, Chen Y. Advanced nitrogen and phosphorus removal in A2O – BAF system treating low carbon to nitrogen ratio domestic wastewater. *rontiers of Environmental Science and Engineering in China*. 2011;5:474–480. Available from: [10.1007/s11783-011-0360-0](https://doi.org/10.1007/s11783-011-0360-0).
- Chen Y, Peng C, Wang J, Ye L, Zhang L, Peng Y. Effects of nitrate recycling ratio on simultaneous biological nutrient removal in a novel anaerobic/anoxic/oxic (A2/O) – biological aerated filter (BAF) system. *Bioresource Technology*. 2011;102:5722–5727. Available from: [10.1016/j.biortech.2011.02.114](https://doi.org/10.1016/j.biortech.2011.02.114).
- Chen Y, Li B, Ye L, Peng Y. The combined effects of COD/N ratio and nitrate recycling ratio on nitrogen and phosphorus removal in anaerobic/anoxic/aerobic (A2/O) – biologi-

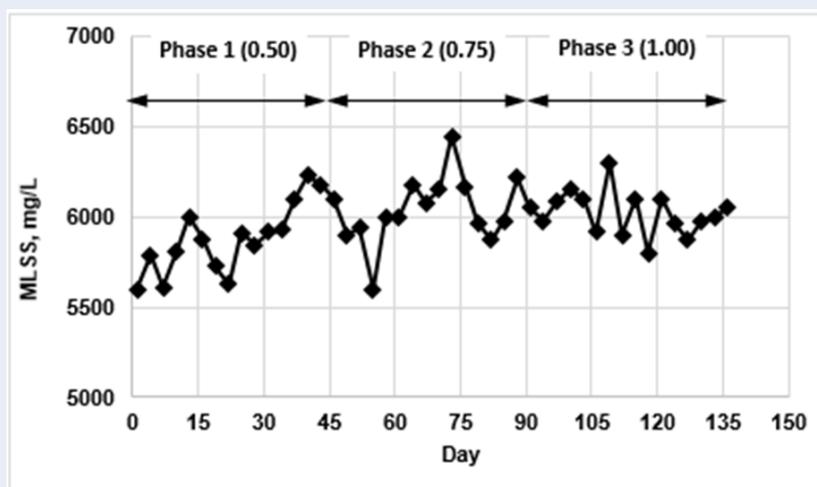


Figure 10: Variation of MLSS concentration during the operational period.

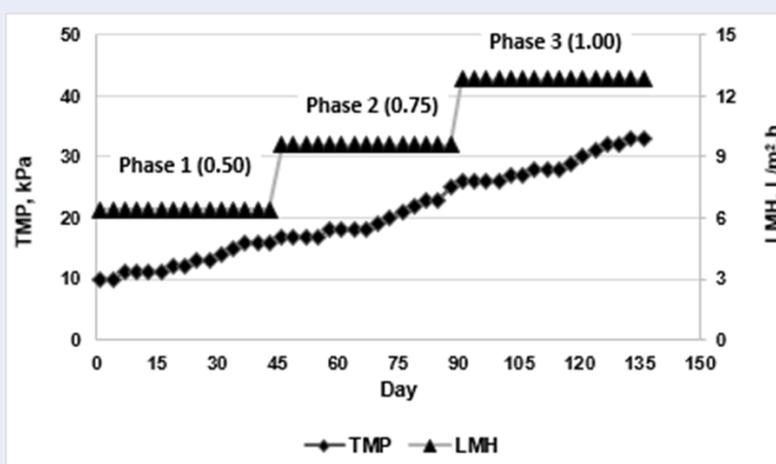


Figure 11: Variation of TMP during the operational period.

- cal aerated filter (BAF) systems. *Biochemical Engineering*. 2015;93:235–242. Available from: [10.1016/j.bej.2014.10.005](https://doi.org/10.1016/j.bej.2014.10.005).
11. Zhang W, Peng Y, Ren N, Liu Q, Chen Y. Improvement of nutrient removal by optimizing the volume ratio of anoxic to aerobic zone in AAO – BAF system. *Chemosphere*. 2013;93(11):2859–2863. Available from: [10.1016/j.chemosphere.2013.08.047](https://doi.org/10.1016/j.chemosphere.2013.08.047).
  12. Judd S, Judd C. *The MBR Book: principles and applications of membrane bioreactors in water and wastewater treatment*. Butterworth-Heinemann Publishing; 2006. p. 209–288. 2nd edition.
  13. Arévalo J, Ruiz LM, Parada-Albarracín JA, González-Pérez DM, Moreno JPB, Gómez MA. Wastewater reuse after treatment by MBR. Microfiltration or ultrafiltration? *Desalination*. 2012;299:22–27. Available from: [10.1016/j.desal.2012.05.008](https://doi.org/10.1016/j.desal.2012.05.008).
  14. Barreto CM, Garcia HA, Hooijmans CM, Herrera A, Brdjanovic D. Assessing the performance of an MBR operated at high biomass concentrations. *International Biodeterioration & Biodegradation*. 2017;119:528–537. Available from: [10.1016/j.ibiod.2016.10.006](https://doi.org/10.1016/j.ibiod.2016.10.006).
  15. Meng F, Zhang S, Oh Y, Zhou Z, Shin HS, Chae SR. Hang-Sik Shin, So-Ryong Chae. Fouling in membrane bioreactors: An updated review. *Water Research*. 2017;114:151–180. Available from: [10.1016/j.watres.2017.02.006](https://doi.org/10.1016/j.watres.2017.02.006).
  16. Banu JR, Uan DK, Yeom IT. Nutrient removal in an A2O-MBR reactor with sludge reduction. *Bioresource Technology*. 2009;100(16):3820–3824. Available from: [10.1016/j.biortech.2008.12.054](https://doi.org/10.1016/j.biortech.2008.12.054).
  17. Falahti-Marvast H, Karimi-Jashni A. Performance of simultaneous organic and nutrient removal in a pilot scale anaerobic–anoxic–oxic membrane bioreactor system treating municipal wastewater with a high nutrient mass ratio. *International Biodeterioration & Biodegradation*. 2015;104:363–370. Available from: [10.1016/j.ibiod.2015.07.001](https://doi.org/10.1016/j.ibiod.2015.07.001).
  18. Khan SJ, Ilyas S, Javid S, Visvanathan C, Jegatheesan V. Performance of suspended and attached growth MBR systems in treating high strength synthetic wastewater. *Bioresource Technology*. 2011;102:5331–5336. Available from: [10.1016/j.biortech.2010.09.100](https://doi.org/10.1016/j.biortech.2010.09.100).
  19. Clescerl LS, Greenberg AE, Eaton AD. *Standard Methods for the Examination of Water and Wastewater*. APHA Publishing;

- 1998.
20. Mateju V, Cizinska S, Krejci J, Janoch T. Biological water denitrification. A review. *Enzyme and Microbial Technology*. 1992;14(3):170–183. Available from: [10.1016/0141-0229\(92\)90062-S](https://doi.org/10.1016/0141-0229(92)90062-S).
  21. Sun XMFY, Wang XY, Li. An innovative membrane bioreactor (MBR) system for simultaneous nitrogen and phosphorus removal. *Process Biochemistry*. 2013;48:1749–1756. Available from: [10.1016/j.procbio.2013.08.009](https://doi.org/10.1016/j.procbio.2013.08.009).
  22. Yun Sun F, Mao Wang X, Yan Li X. An innovative membrane bioreactor (MBR) system for simultaneous nitrogen and phosphorus removal. *Bioresource Technology*. 2009;100:1048–1054. Available from: [10.1016/j.biortech.2008.07.045](https://doi.org/10.1016/j.biortech.2008.07.045).
  23. Meng F, Chae SR, Drews A, Kraume M, Shin HS, Fenglin Yang. Recent advances in membrane bioreactors (MBRs): Membrane fouling and membrane material. *Water Research*. 2009;43(6):1489–1512. Available from: [10.1016/j.watres.2008.12.044](https://doi.org/10.1016/j.watres.2008.12.044).

# Nghiên cứu hệ thống A<sub>2</sub>O-MBR để loại bỏ đồng thời thành phần nitơ và photpho có trong nước thải sản xuất bia

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## TÓM TẮT

Hệ thống xử lý nước thải sử dụng phương pháp sinh học kỵ khí, thiếu khí, hiếu khí đồng thời và kết hợp màng siêu lọc (A<sub>2</sub>O-MBR) đã được sử dụng nhằm tăng cường khả năng xử lý cùng lúc thành phần nitơ và photpho có trong nước thải sản xuất bia. Cụm A<sub>2</sub>O chứa vi sinh vật với thời gian lưu bùn ngắn chịu trách nhiệm loại bỏ hữu cơ, photpho và khử nitrat. Bể MBR chứa vi sinh vật với thời gian lưu bùn dài chịu trách nhiệm nitrat hóa amoni và tuần hoàn nitrat. Mô hình A<sub>2</sub>O-MBR được làm bằng mica với thể tích chứa 49,5 lít đã được vận hành với thời gian lưu nước giảm dần từ 24, 18 đến 12 giờ tương ứng với tải trọng hữu cơ tăng dần từ 0,50; 0,75 đến 1,00kgCOD/m<sup>3</sup>.ngày. Kết quả cho thấy mô hình không chỉ xử lý tốt thành phần hữu cơ mà còn loại bỏ gần như hoàn toàn nitơ và photpho. Ở cả 03 tải trọng, nồng độ COD đã giảm đáng kể ở các ngăn kỵ khí và thiếu khí của cụm A<sub>2</sub>O, cho thấy phần lớn cơ chất đã được sử dụng ở các ngăn kỵ khí và thiếu khí, tương ứng với các quá trình tách photpho và khử nitrat. Nitrat hóa trong bể MBR đã xảy ra gần như hoàn toàn, với hiệu quả loại bỏ amoni là trên 98%. Khử nitrat trong ngăn thiếu khí được thực hiện nhiều nhất có thể. Các điều kiện cần thiết cho sự phát triển của vi sinh vật tích lũy photpho (PAOs) để loại bỏ photpho bởi tăng cường sinh học (EBPR) có thể đã được cung cấp. Ở tải trọng 0,75kgCOD/m<sup>3</sup>.ngày, hiệu quả xử lý COD, NH<sub>4</sub><sup>+</sup>-N, TN và TP của mô hình là cao nhất, tương ứng với 95,4; 99,2; 86,7 và 84,6%. Giá trị đầu ra của các thông số này là nằm trong giới hạn cho phép của QCVN 40:2011/BTNMT, cột A. Mô hình A<sub>2</sub>O-MBR có khả năng đạt được đầu ra của nước thải sản xuất bia với nồng độ nitơ và photpho rất thấp.

**Từ khóa:** Hệ thống A<sub>2</sub>O-MBR, nước thải sản xuất bia, loại bỏ nitơ, loại bỏ photpho

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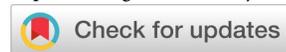
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## Lịch sử

- Ngày nhận: 18-1-2019
- Ngày chấp nhận: 14-5-2019
- Ngày đăng: 20-6-2019

## DOI:

<https://doi.org/10.32508/stdjsee.v3i1.507>



## Bản quyền

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**Trích dẫn bài báo này:** Thiên V N T, Hùng D V, Hoa N T T. **Nghiên cứu hệ thống A<sub>2</sub>O-MBR để loại bỏ đồng thời thành phần nitơ và photpho có trong nước thải sản xuất bia**. *Sci. Tech. Dev. J. - Sci. Earth Environ.*; 3(1):12-22.